# Incorporating connectivity into conservation planning for the optimal representation of multiple species and ecosystem services

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**Abstract:** Conservation planning tends to focus on protecting species' ranges or landscape connectivity but seldom both—particularly in the case of diverse taxonomic assemblages and multiple planning goals. Therefore, information on potential trade-offs between maintaining landscape connectivity and achieving other conservation objectives is lacking. We developed an optimization approach to prioritize the maximal protection of species' ranges, ecosystem types, and forest carbon stocks, while also including habitat connectivity for range-shifting species and dispersal corridors to link protected area. We applied our approach to Sabah, Malaysia, where the state government mandated an increase in protected-area coverage of approximately 305,000 ha but did not specify where new protected areas should be. Compared with a conservation planning approach that did not incorporate the 2 connectivity features, our approach increased the protection of dispersal corridors and elevational connectivity by 13% and 21%, respectively. Coverage of vertebrate and plant species' ranges and forest carbon and 3% less butterfly range than when connectivity features were not included. Hence, the inclusion of connectivity into conservation planning can generate large increases in the protection of landscape connectivity with minimal loss of representation of other conservation targets.

**Keywords:** Borneo, climate change, connectivity, corridors, deforestation, habitat loss, rainforest, systematic conservation planning

Incorporación de la Conectividad a la Planeación de la Conservación para la Representación Óptima de Especies Múltiples y Servicios Ambientales

**Resumen:** Las tendencias de planeación de la conservación tienden a enfocarse en la protección de la distribución geográfica de las especies o en la conectividad de paisajes, pero rara vez se enfocan en ambas – particularmente

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934

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para el caso de los ensamblajes taxonómicos y las metas múltiples de planeación. Por lo tanto, hay carencias en la información sobre las compensaciones potenciales entre mantener la conectividad de los paisajes y alcanzar otros objetivos de conservación. Desarrollamos una estrategia de optimización para priorizar la protección máxima de la distribución de las especies, los tipos de ecosistemas y los stocks de carbono de los bosques, a la vez que incluimos la conectividad del hábitat para las especies que modifican su distribución y los corredores de dispersión para conectar el área protegida. Aplicamos nuestra estrategia en Sabah, Malasia, en donde el gobierno estatal ordenó un incremento de ~305, 000 ha en la cobertura de áreas protegidas sin especificar la ubicación de las nuevas áreas protegidas. En comparación con una estrategia de planeación de la conservación que no incorporó las dos características de la conectividad, nuestra estrategia incrementó la protección de los corredores de dispersión y la conectividad altitudinal en un 13% y 21% respectivamente. La cobertura de la distribución de las especies de plantas y vertebrados y de los tipos de bosque fue la misma con o sin la inclusión de la conectividad. Nuestra estrategia protegió 2% menos del carbono forestal y 3% menos de la distribución de mariposas que cuando no se incluyeron las características de conectividad en la estrategia. Por lo tanto, incluir a la conectividad en la planeación de la conservación puede generar grandes incrementos en la protección de la conectividad del paisaje con una pérdida mínima de representación para los demás objetivos de conservación.

**Palabras Clave:** Borneo, cambio climático, conectividades, corredores, deforestación, pérdida de hábitat, planeación sistemática de la conservación, selva

# Introduction

Protected areas are critical but insufficient for mitigating the impacts of habitat loss, which is the principal threat to global biodiversity (Laurance et al. 2012; Gray et al. 2016). In 2010, signatories to the Convention on Biology Diversity adopted a commitment to protect 17% of global land and inland sea surface area by 2020 (known as Aichi Biodiversity Target 11). As of July 2018, the coverage of terrestrial protected areas had reached 14.9%, which implies that achieving the target requires an additional approximately 2.8 million km<sup>2</sup> of protected areas globally by 2020. Locating new protected areas in the lowland tropics, and especially in Southeast Asia, would contribute disproportionately to global biodiversity conservation through avoided deforestation in biodiversity hotspots (Myers et al. 2000).

Determining the most important areas for additional conservation measures requires prioritizing multiple objectives. Many conservation plans seek to maximize representation of species' ranges (Pressey et al. 2007). Indeed, powerful and widely used prioritization tools such as Marxan (Ball & Possingham 2000) and Zonation (Moilanen 2007) have helped planners optimize the selection of new conservation areas on every continent (Sinclair et al. 2018). Other prioritization plans have explicitly focused on landscape connectivity (e.g., Gordon et al. 2009; Lehtomäki et al. 2009; Sirkiä et al. 2012) because this affects metapopulation persistence (Fahrig & Merriam 1994; Hanski & Ovaskainen 2000), supply of ecosystem services (Kukkala & Moilanen 2017), and organisms' adaptive capacities in the face of climate change (Scriven et al. 2015; Reside et al. 2017). The technical capacity for assessing connectivity and determining optimal locations for habitat corridors is growing rapidly (e.g., Lehtomäki & Moilanen 2013; Brodie et al. 2016; Daigle et al. 2018).

Far fewer conservation plans, however, simultaneously optimize landscape connectivity and the representation of multiple conservation features (Reside et al. 2017; Harlio et al. 2019), particularly for multispecies assemblages, over large landscapes, and in the context of diverse planning goals (Magris et al. 2018). Simultaneously accounting for both representation of conservation features and connectivity increased the predicted average population size of Mediterranean fishes by approximately 66% (Magris et al. 2018) and the average metapopulation capacity (a proxy for population persistence) of Australasian mammals by 5-fold (Strimas-Mackey & Brodie 2018). But jointly optimizing connectivity and representation has seldom been attempted with more diverse taxonomic assemblages and multiple conservation objectives (e.g., species' ranges, ecosystem services, and habitat types).

Moreover, existing approaches to incorporating connectivity into conservation planning tend to focus on single aspects of connectivity. For example, corridors may be identified so as to maintain dispersal among habitat patches and thereby enhance metapopulation stability (Fahrig & Merriam 1994; Hanski & Ovaskainen 2000) or to provide movement routes by which organisms can shift their distributions in response to climate change (Scriven et al. 2015; Reside et al. 2017). However, simultaneously incorporating the different facets of connectivity into planning has been done much more rarely (Moilanen & Wintle 2007; Lehtomäki et al. 2009), despite each facet being critical to long-term persistence of species in dynamic landscapes.

Multifaceted conservation optimization always has the potential to involve trade-offs. Trying to protect one conservation feature can decrease protection of another feature that is uncorrelated or negatively correlated with the first. The inclusion of connectivity in spatial prioritization involved relatively small trade-offs with habitat quality in a semiarid grassland site in Europe (Harlio et al. 2019) and with estimated current and future species' ranges of tropical vertebrates in Australia (Reside et al. 2017). But there is little understanding of connectivity trade-offs in prioritization analyses involving more diverse taxa and conservation objectives. For example, it remains important to assess how the addition of connectivity to multiobjective spatial planning in complex systems would affect the areal representation of species' ranges, habitat types, and ecosystem services. The lack of evaluations of connectivity in the context of complex, multifaceted spatial conservation planning may have contributed to the only slight improvements in landscape connectivity that have accompanied recent increases in the coverage of protected areas (Saura et al. 2018).

We designed a method to simultaneously prioritize species' ranges, ecosystem services (aboveground forest carbon storage), and 2 types of landscape connectivity. We then assessed whether the inclusion of connectivity compromised the achievement of the other conservation objectives. Our approach explicitly incorporated both the specific locations of dispersal corridors based on animal movement models and connectivity of habitat along elevational gradients to enhance climate change resiliency and facilitate shifts in species' ranges. We applied this prioritization approach to recommend new areas for protection of rainforest in the state of Sabah in Malaysian Borneo, where the state government has mandated an increase in the coverage of terrestrial totally protected areas (TPAs) (e.g., state parks or class 1 forest reserves) of approximately 305,000 ha. The locations of the forest areas to be designated for protection have not yet been chosen. We developed our method to aid the Sabah Forestry Department in this decision making.

## Methods

#### Study System and Objectives

Sabah is in the north of Borneo, which is a global biodiversity hotspot due to its high levels of endemism and rapid land conversion (Myers et al. 2000). Rates of landuse change in this region have been among the fastest in the world (Langner et al. 2007; Miettinen et al. 2011), often driven by the expansion of industrial-scale tree and agricultural (mainly oil palm) plantations (Gaveau et al. 2018). Until the 1970s, most of Borneo was covered by primary rainforest (Bradshaw et al. 2009). By 2010 just over 50% of the island remained forested, much of it commercially selectively logged and fragmented (Gaveau et al. 2014). The Sabah Forest Policy directives call for an increase in the coverage of terrestrial TPAs to 30% by the year 2025 (SFD 2018), which requires a minimum of 304,708 ha of additional protected area. For our conservation planning exercise, we were asked by the Sabah Forestry Department to prioritize 410,000 ha for protection because this would allow decision makers flexibility in choosing the final configuration of new TPAs.

Within the study area, we restricted our analysis to mainland Sabah in areas with forest cover (Fig. 1; Supporting Information). We determined the remaining forested area of Sabah, excluding mangroves, by applying a threshold of 40 Mg/ha of aboveground forest carbon, as determined from a recent high-resolution mapping study (Asner et al. 2018). This followed similar forest delineation guidelines used elsewhere (Rosoman et al. 2017) and was selected to be low enough to include areas of minimally degraded forest capable of regeneration and to exclude most oil palm and short-rotation tree plantations.

We employed a conservation-planning framework with 3 underlying objectives. First, we aimed to ensure that the areas of forest recommended for protection covered distributional ranges for a variety of taxa for which such data were available—plants, butterflies, and several vertebrate groups (amphibians, birds, and mammals). We included only restricted-range or threatened species to ensure that we prioritized the species of highest conservation value. Overall we included 149 range-restricted plants (trees, shrubs, and orchids), 77 range-restricted butterflies, and 83 threatened vertebrates (International Union for Conservation of Nature status of vulnerable, endangered, or critically endangered [Supporting Information]).

Our second objective was to ensure that landscape connectivity was conserved. Landscape connectivity was measured in 2 different ways, reflecting its different ecological benefits. First, we sought to protect forest areas that provide linkage between existing TPAs, specifically allowing for regular population exchange of highly mobile species, for example, clouded leopards (*Neofelis diardt*) (hereafter, dispersal corridors). Second, we aimed to conserve forest areas that spanned elevational gradients between TPAs so as to provide range shifting routes for less mobile species (e.g., a forest-dependent butterfly) in the face of climate change (hereafter, elevational connectivity).

Our final objective was to maximize representation of different forest types (excluding mangroves and beach forest), as a proxy for facets of biodiversity we did not capture with our species' ranges, and to protect forest areas that store particularly large amounts of aboveground carbon, so as to contribute to state and national commitments to climate change mitigation. For forest types, we used data from the Sabah Forest Research Centre comprising a classification of the state's forest types based on the predevelopment spatial distribution of each. The 22 forest types (Supporting Information) were distinguished by species groups, edaphic characteristics, and land formations. Details of the different types of feature layers (species' ranges, dispersal corridors, elevational connectivity, forest type, and aboveground carbon) are provided in the Supporting Information.



Figure 1. Study area (a) on mainland Sabab, Malaysian Borneo (dark green, existing protected areas; light green, other remaining forested areas) areas prioritized for additional protection (orange and blue) (b) with and (c) without explicit incorporation of dispersal corridors and elevational connectivity, and (d) overlap of prioritized areas in (b) and (c) (red).

#### **Estimation of Connectivity**

To determine the location of dispersal corridors, we simulated the movements of individuals across the landscape using correlated random walk models to identify forest areas that may be the most important for animal movement between existing TPAs. Within these simulations, we incorporated biologically relevant parameters to generate movement scenarios commonly associated with transit or natal dispersal as exhibited by wide-ranging mammals and birds. We varied starting locations within areas of high-quality forest habitat as well as the resistance to movement through suboptimal habitat (nonforested areas). This technique allowed the estimation of the flow of dispersers across all potential corridors on the landscape, or the centrality of each linkage (Brodie et al. 2016). We weighted each landscape unit according to its spatial position relative to its 2 nearest TPAs as well as the size of the TPAs. Previous research in the system suggests that short corridors linking 2 large habitat patches are the most important to metapopulation persistence (Brodie et al. 2016). The output value from the movement simulations was the total use of each cell in the landscape by all dispersing individuals that successfully reached a new habitat patch. The most important corridors were then generated by iteratively selecting the highest value cells across the landscape until a corridor across the landscape was reached. By integrating animal movement and location weighting, we generated a connectivity metric that incorporated movement stochasticity and metapopulation dynamics to identify corridors of contiguous planning units (details in Supporting Information).

To assess elevational connectivity, we used Condatis models (Hodgson et al. 2012), based on circuit and metapopulation theories, to identify the most important forested areas connecting each lowland (source) TPA to higher elevation (target) TPA grid cells: those that would have equal or lower mean annual temperature in 2061–2080 (based on Relative Concentration Pathway 8.5, the business-as-usual emissions scenario) than the source TPA's current mean annual temperature. Identifying the most important forest areas was achieved by progressively dropping unprotected forest from the landscape (Hodgson et al. 2016) and monitoring the decline in the conductance between the source and target TPAs, which indicated the rate at which organisms could cross the landscape (details in Supporting Information).

#### **Conservation Prioritization**

We applied systematic conservation planning by using integer linear programing based on specified objectives, input conservation features, targets, and land budget. This was implemented with the package priotritizr 4.0.2 (Hanson et al. 2017) in the program R (R Core Team 2015). Because our aim was to incorporate several types of conservation features, each with varying numbers of raw feature layers, we developed a 2-step prioritization process (Supporting Information). First, we prioritized the input features for each of 7 categories (i.e., plants, butterflies, vertebrates, aboveground carbon, forest types, elevational connectivity, and dispersal corridors) that addressed our 3 ecological objectives and our budgeted land area with the objective of maximizing the number of features that met a specified target without exceeding the land-area budget. This produced a single prioritized output layer per feature category (Supporting Information). Second, we used these 7 feature category layers (e.g., a single prioritized layer for butterflies instead of 77 individual species layers), all weighted equally and with equal representation targets, as input features to create one overall output layer. We used a boundary length modifier (Supporting Information) to clump the prioritized areas spatially. Existing TPAs were ensured of being included in the conservation solution. Thus, our solution found the optimal area (up to 410,000 ha) outside of existing TPAs that maximized the representation of species' ranges, forest type diversity, aboveground carbon density, dispersal corridors, and elevational connectivity across Sabah. Prioritization was implemented across 100-ha planning units covering the current extent of forested areas in the state.

#### **Evaluation of the Conservation Prioritization**

We computed the total proportion of each feature layer that was covered by the prioritized area, both excluding and including existing TPAs. For our first assessment, excluding existing TPAs and evaluating only the additional area prioritized for conservation, the 410,000 ha land budget accounted for approximately 5% of the total land area of mainland Sabah. Thus, we measured the success of the 410,000 ha of prioritized new areas by assessing the number of raw feature layers for which the overall prioritization solution reached a coverage of at least 5%. It is also important to consider that many of the best areas are already protected by the existing TPA network, so reaching this level of coverage for each of the raw features is more difficult in the remaining available forested area of Sabah.

For our second assessment, the total area of our full prioritized solution (i.e., our land budget plus existing TPAs) was 32% of the total land area of mainland Sabah. Thus, to measure the efficacy of our prioritization analysis, we determined how many raw input features reached at least 32% coverage through the full prioritization solution.

We determined how the results of our overall prioritization compared with those generated from a more typical systematic conservation planning exercise that did not explicitly incorporate connectivity. As a measure of this more typical analysis, we ran the same prioritization analysis using the 5 feature layers not related to connectivity (i.e., plants, butterflies, vertebrates, forest types, and aboveground carbon) but excluded dispersal corridors and elevational connectivity. Finally, we assessed the sensitivity of our overall prioritization results to each input feature layer (Supporting Information).

# Results

The total area available for conservation in our analysis covered 2,288,426 ha, of which 1,879,238 ha were in existing TPAs on mainland Sabah that were forced to be included in the solution. Our 2-step analysis prioritized an additional 409,187 ha for conservation (hereafter, prioritized area). The prioritized area (Fig. 1b) was concentrated largely in several clumps including southwestern Sabah, which had high species richness of butterflies and vertebrates and important dispersal corridor connections to Sarawak and Kalimantan, the Upper Kinabatangan, which was important for plants, vertebrates, and elevational connectivity, and areas east of the Crocker Range that were important for all taxa and contained a diverse range of forest types (Supporting Information). All these areas were also relatively high in aboveground carbon density, generally above 200 Mg C/ha (Supporting Information).

In our assessment of how well the 410,000 ha prioritized area covered each of the raw input conservation features outside of existing TPAs, 200 out of the 246 raw input features (excluding the restricted plants species that had limited point locality data) reached the  $\geq$ 5% benchmark. On average, raw input features reached over 12% coverage (Fig. 2), demonstrating that our prioritization analysis provided a conservation solution that was ~2.4 times better than what would have been achieved by randomly selecting new conservation areas. Connectivity features had the lowest coverage of the 7 input features of the final prioritization. These features were



generated by analyses that focused explicitly on locations outside the existing TPA network, so existing TPAs added no coverage for these features.

For our assessment of how well the full conservation solution (i.e., the prioritized area plus existing TPAs) covered the raw input features, an average of 37% coverage was reached for species' ranges and forest types (Table 1). The mean coverage for all 7 feature categories was  $\geq 40\%$ (Fig. 2). Coverage was achieved most successfully for butterflies and forest types (58% of the input features for each). Coverage was least successful for vertebrates (43% of input features), though even for this feature, the fact that coverage was so far above the 32% benchmark suggests, again, that the prioritization analysis provided a substantially better solution than randomly choosing conservation areas would have done.

Compared with the analysis that excluded the 2 connectivity features (i.e., a more typical systematic conservation planning approach), our prioritized area showed no difference in protected coverage of vertebrates, plants, and forest types, 3% less coverage for butterflies, Figure 2. Proportion of the conservation features (input feature data layers) protected by the overall conservation solution (a) with and (b) without explicit incorporation of dispersal corridors and elevational connectivity (middle borizontal line, median; upper and lower box ends, 75th and 25th percentiles, respectively; upper and lower vertical lines [wbiskers], largest and smallest values not further than 1.5 times the inner quartile range; dots, data points).

2% less coverage for aboveground carbon, 12.5% more coverage of elevational connectivity, and 21.4% more protection of dispersal corridors (Table 1).

# Discussion

Systematic conservation planning is important for identifying the most important areas for new protection in order to stem extinctions driven by habitat loss. The Convention on Biological Diversity's Aichi Target 11 calls for at least 17% of the world's land to be covered by well-connected protected areas or other effective areabased conservation measures by 2020 (CBD 2010). Globally, protected area coverage is approximately 14.7%, but only about half of that area is considered wellconnected (Saura et al. 2018). But although the goal of numerous conservation planning exercises was to optimize the representation of conservation features or the connectivity of the landscape, few sought to do both, especially in the context of multispecies assemblages

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Input feature categories	No. of raw input features for round 1 of prioritization	<i>Targets for round 1</i> of prioritization (%)	Proportion of raw input features protected by overall prioritized solution			Number of raw input
			mean	range	total	coverage
Vertebrates	81	43.5	0.39 (0.40)	0.00-1.00 (0.00-1.00)	-	34 (36)
Invertebrates	77	43.5	0.32 (0.32)	0.25-0.42 (0.25-0.43)	-	35 (37)
Plants	66*	43.5	0.39 (0.39)	0.12-0.88 (0.11-0.89)	-	33 (33)
Forest types	19	50	0.44 (0.45)	0.00-0.94 (0.00-0.96)	-	11 (11)
Aboveground carbon	1	52.5	-	-	0.49 (0.50)	-
Elevational connectivity	1	31	-	-	0.17 (0.14)	-
Dispersal corridors	1	99	-	-	0.17 (0.14)	-

Table 1. Protection targets and achieved protection coverage of raw input features of the overall prioritized solution (existing totally protected areas plus new areas prioritized for conservation) with (no parentheses) and without (parentheses) explicit incorporation of dispersal corridors and elevational connectivity.

\*An additional 83 plant species (331 locality records) with extremely limited distributions were included in this category, represented as 0.05-ba localities that were either included or excluded in their entireties in the prioritized area.

and diverse conservation objectives. Our prioritization of multiple conservation features (ranges for a diverse array of species, ecosystem types, and forest carbon storage) across a large landscape and incorporation of landscape connectivity greatly enhanced the protection of dispersal corridors and elevational gradients along which species could move in response to warming temperatures, with no additional cost in terms of land area requirements, and limited costs in terms of poorer representation of other conservation features. This is consistent with other studies that had less diverse conservation objectives and showed relatively limited trade-offs between the protection of connectivity and species' ranges (Reside et al. 2017; Harlio et al. 2019). Part of the reason that these studies and ours may have been able to achieve multiple benefits with limited costs is that at least some of the input features were spatially correlated (Supporting Information). At least in our case, this is unlikely to represent systematic collection bias because we included a wide range of taxa representing data assembled from multiple sources as input features. Applications of these methods to other systems with less input feature correlation, for example, where species were strongly clustered into distinct, nonoverlapping habitat types, could increase the difficulty in simultaneously achieving different conservation objectives.

The locations selected for conservation generally contained several conservation features, but full representation of all 7 conservation features was only achieved across the suite of areas. Some of the hyperendemic plants (where we used point locality information instead of predicted species' ranges) were not included in our prioritized area (Table 1), and may require additional conservation measures in particular localities. The most important part of the state for dispersal corridors was southwestern Sabah, where connectivity is needed to link the cluster of protected areas in the central part of Sabah to protected forests in Sarawak and Indonesia (Brodie et al. 2016). Several areas were critical for elevational connectivity, including the Upper Kinabatangan in central Sabah and areas around Kinabalu Park that would increase its connection to the lowlands. The Kinabalu area was also identified as a priority area in a conservation analysis focused on mammals (Struebig et al. 2015). These 3 areas were also rich in vertebrates (southwest Sabah, Deramakot), butterflies (southwest Sabah), plants (Deramakot, Kinabalu Park vicinity), forest types (southwest Sabah, Kinabalu Park vicinity), and aboveground carbon (southwest Sabah, Deramakot) (Supporting Information).

In Sabah the protection of the prioritized area identified by our analysis would provide important conservation benefits. The prioritized area has been presented to the Sabah Forestry Department, which has initiated a round of free, prior informed consent consultations with indigenous peoples, local communities, and other stakeholders who may have overlapping rights and interests in the locations identified for enhanced conservation. Outcomes of this process will inform subsequent revisions of the prioritization analysis. We will continue with this iterative process to seek consensus on the final recommendation for the designation of new protected areas. The conservation influence of this and other planning measures could be enhanced further via cross-border coordination with governments in Brunei Darussalam, Kalimantan (Indonesia), and Sarawak (Malaysia) (van Paddenburg et al. 2012; Runting et al. 2015).

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## **Supporting Information**

Details of the input feature data for the conservation prioritization analysis (Appendix S1), species of plants, vertebrates, and butterflies included in the prioritization (Appendix S2), forest types included in the prioritization (Appendix S3), and supplemental methods and results figures (Appendix S4) are available online. The authors are solely responsible for the content and functionality of these materials. Queries (other than absence of the material) should be directed to the corresponding author. The analysis and prioritization code is available at github.com/williash23/SEARRP-UMT-Prioritization.

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